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SPATIAL STRUCTURE AND CO_2 Emissions Due to Commuting: An Analysis on Italian Urban Areas

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Spatial Structure and CO₂ Emissions Due to Commuting: an Analysis on Italian Urban Areas^{*}

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Abstract

The aim of this paper is to investigate whether and to what extent the spatial configuration of an urban area affects its level of environmental externalities. Starting from previous contributions to this field of research, it examines several features of urban spatial structure – such as compactness, monocentricity, concentration and functional diversity – and attempts to gauge their environmental implications in terms of per capita CO_2 emissions associated with a given pattern of commuting (i.e., mode of commuting and distance travelled). The main finding of the analysis on the 111 largest Italian urban areas is that urban spatial configuration is an important determinant of travel patterns and the associated level of per capita CO_2 emissions. In particular, smaller, more compact and less monocentric areas are associated with lower levels of CO_2 per commuter, with sociodemographic characteristics also playing a role.

Key words: urban spatial structure, commuting, environmental costs, CO₂ emissions.

JEL classification codes: Q56, R14, R41.

1. INTRODUCTION

Daily traffic flows are at the origin of many environmental externalities generated by economic and social processes. Among such externalities, air pollution and climate change have attracted much attention among researchers and policymakers, provoking a controversy about the most effective ways to reduce these negative environmental effects. In this regard, some scholars – including new urbanists and advocates of "smart

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growth" – argue that the characteristics of built environment and of urban spatial structure play an important role in shaping the patterns of mobility.

One assumption often made is that a low-density and dispersed urban area (UA) tends to be more dependent upon automobile use and requires longer commutes, proving to be less sustainable in terms of air pollution and global warming externalities. However, It is debatable whether a compact area performs better than a dispersed one in terms of CO_2 and other pollutant emissions. In fact, while it is widely recognised that the latter model of urban development encourages a more intense use of private motorised means of transport, it is also maintained that the former is likely to be beset with congestion problems – which make journeys-to-work slower – thus increasing polluting emissions per kilometre travelled. Moreover, the advocates of the *ville emergent* have argued that free-market decisions about localisation and construction – which is more likely to lead to scattered development – could be more environmentally sustainable than centralised planning (Dubois-Taine and Chalas, 1997).

A model of "compact" urban development may prove less efficient also given individuals' preference to live in low-density suburbs, which tend to be greener and safer than dense central areas in the US (Bruegmann, 2005). This point may not be relevant to the European context, where the hinterlands of metropolitan areas are rarely characterised by improved urban quality (Calafati, 2008). Indeed, even assuming that a scattered spatial structure is associated with commuting externalities, the latter could be reduced by means of Pigouvian tools, such as taxes and congestion tolls. Solutions of this kind, however, more common in the US, appear hardly feasible in Europe, where taxation on traffic and especially on fuel is already high.

The aim of this paper is to verify whether and to what extent the characteristics of urban spatial structure explain the levels of CO_2 emissions generated by commutes. In order to answer this research question, an indicator of per commuter CO_2 emissions is proposed on the basis of the characteristics of home-work trips within each UA (i.e., transport mode and distance travelled). Secondly, the spatial configuration of the UAs is examined by looking at various spatial dimensions, such as the degree of compactness of the residential structure, the archetypical model of spatial development (monocentric, polycentric or dispersed), the degree of clustering of activities and its internal functional diversity. In this way it is

possible to isolate the role of different elements that characterise contemporary UAs, instead of focusing on residential density alone.

The research question outlined above is addressed through an empirical analysis on the Italian urban system. Nationwide studies on this issue are still lacking in Italy, although the peculiar spatial development that has characterised Italy in the last decades (Calafati, 2009) would warrant specific attention. Nonetheless, the scientific community has significantly contributed to this field of research with a number of case studies (Camagni *et al.* 2002b, Travisi *et al.*, 2010). In this work the empirical analysis is carried out on the 111 largest Italian UAs, identified with those Local Labour Systems (LLSs)¹ whose central municipality had at least 50 000 inhabitants in 2001. The LLS has been chosen as the unit of analysis because it makes it possible to focus on clusters of integrated municipalities rather than just on single urban municipalities, given the profound reorganisation of the Italian territory in recent decades (see Section 2).

The work is structured as follows: Section 2 looks at the evolution of the spatial structure of the Italian UAs in recent decades and briefly reviews its main implications in terms of collective – and especially environmental – costs. Section 3 examines the relationship between UA spatial configuration and its internal pattern of commutes against an appropriate theoretical background. Section 4 provides a descriptive analysis of both urban spatial structures and commuting patterns in Italy, while Section 5 introduces the CO_2 environmental indicator and the other relevant variables. Section 6 presents the empirical analysis and its results, which are commented upon and questioned. Finally, Section 7 makes some concluding remarks and puts forward some policy recommendations.

2. THE CHANGING SPATIAL CONFIGURATION OF ITALIAN URBAN AREAS

The archetypical Italian (and European) city exhibits the well-known characteristics of the *compact city*. It is densely inhabited, with a strong urban identity and a high "relational density" based on proximity (Burton, 2000). During the past five decades, however, a new pattern of urban

¹ A LLS is defined as a cluster of contiguous municipalities with a high degree of selfcontainment in terms of journeys-to-work, thus constituting a single labour market (ISTAT-Sforzi, 1997).

development has arisen in Italy too which has led to the formation of the socalled *dispersed city*. Today, all the contemporary Italian UAs can be considered dispersed cities (Secchi, 2005).

Albeit belatedly if compared with the US and other European countries, the Italian territory has undergone a "restructuring process" that has given rise to new forms of spatial organisation. Following a period extending approximately from the 1950s to the 1970s – of intense spatial concentration of both population and employment, the major Italian cities have experienced a process of suburbanisation and decentralisation that has gradually led to the formation of many dispersed UAs (Cirilli, 2010). In most of the latter, the municipalities surrounding the central cities have increased the degree of their physical and relational integration. This phenomenon of "territorial coalescence" (Calafati, 2009) has often resulted in rapid and disordered expansion of settlements, thus obfuscating the historical polycentricity of European UAs. In the 1990s, European policymakers began to address the costs associated with urban dispersion, and a wide-ranging debate arose among scholars on the economic and social rationale of this pattern of spatial organisation (Bruegmann, 2005; Burchell et al. 2005; Brueckner, 2001; Altshuler, 1997; Gordon and Richardson, 1997).

2.1 Urban dispersion and its determinants

Urban dispersion can be defined as a pattern of spatial development characterised by a low residential density, a small concentration of population and employment in the central business district (CBD), and a low degree of proximity and functional diversification (Muñiz *et al.*, 2006). Physical discontinuity may be another of its distinguishing features, in that newly-built settlements are not necessarily contiguous to the older city.

Urban dispersion has evidently been fostered by the fast-growing rate of new house construction in recent decades. This in its turn has been determined by a wide array of factors, ranging from decreasing interest rates to the rising number of one-person families and "second homes?" (Trilla, 2001). In regard to the localisation of urban activities, residential functions have been crowded out in many UAs' CBD by high value-added tertiary activities able to afford higher rents. Furthermore, the revitalisation of existing real estate in historic city centres may be much costlier than building new settlements in suburban areas. A dispersed (and disordered) urban development may also be the undesirable outcome of a high degree of administrative fragmentation and of fiscal competition among local municipalities to attract investments. Likewise, the lack of effective public decision-making at the appropriate territorial scale has often resulted in the building of isolated agglomerations of houses, thus leading to the formation of discontinuous and highly land-consuming settlements.

Urban dispersion is also the result of wider changes in terms of technological progress and individual preferences. Advances in the ICT industry have significantly influenced localisation decisions. Enhanced centralised control and coordination in large firms' headquarters, for instance, have favoured the territorial dispersion of their productive units and back-office activities (Sassen, 2006). Moreover, technological progress in the transport sector, as well as public programmes for road infrastructure construction, have decreased transportation costs, hence widening the radius of *circadian cycles* without increasing travel duration. In regard to the role of individual preferences, an increasing number of households have come to prefer low-density areas, which allow closer contact with nature, a quieter environment and lower congestion levels (Gordon and Richardson, 1997). In addition, isolated housing permits a higher degree of social segregation based on income (and cultural) differences (Camagni *et al.*, 2002a).

2.2 The effects of urban dispersion

The advocates of public intervention in terms of urban planning argue that free interactions among individual preferences are unlikely to lead to an efficient outcome because of several market failures (Banister, 1997; Brueckner, 2001; Camagni *et al.*, 2002b). The latter ultimately stem from economic as well as social and environmental costs incurred by society, but individual agents do not fully (or even at all) take these into account when they make their decisions regarding transport mode, localisation and land use. From this perspective, urban dispersion is often associated with a variety of "collective" costs (Altshuler, 1997; Burchell *et al.*, 2001; Calafati, 2003).

As regards pure economic costs, a recent study applied to a Spanish case has estimated that private costs – costs of projecting, licensing and building, other urbanisation burdens, heating, water, cleaning and power consumption – and public costs – connection to water and sewage systems and to other utilities – are, respectively, two and seven times higher for an isolated house than a flat in a block (Henry, 2007).

Urban dispersion also entails major non-economic costs (Massey and Denton, 1988; 1993; Deurloo and Musterd, 1998; Cutler *et al.*, 1999). Firstly, a dispersed UA tends to be less accessible, especially to children, the elderly and the disabled. Secondly, there is a tendency for a clear separation to arise among different social groups because of income levels or racial identity. Thirdly, high residential densities may enhance the sense of safety (Jacobs, 1961; Elkin *et al.*, 1991), although a positive relationship between density and urban crime has been found as well (Newman, 1972; Coleman, 1985; Burton, 2000). In a dispersed area, moreover, urban identity tends to weaken, though it may turn into a "sense of community" (Delgado, 1999). This effect is often associated with scarce acceptance of novelty, diversity and tolerance, and thus – following Florida's theory (2002) – with less capacity to innovate.

Two groups of effects can be identified in regard to the environmental implications of urban dispersion. One concerns the direct environmental costs of house building undertaken in a discontinuous or scattered manner, as opposed to a more compact pattern of development. Isolated housing implies more waterproof land and water consumption, a larger loss of rich soil and a higher degree of land fragmentation – factors which in their turn entail less diversity in land uses. The other effect stems from the pattern of commuting associated with urban dispersion (Anderson *et al.*, 1996) and consists in noise and air pollution, occupation of land potentially available for more ecological uses and, finally, traffic accidents. Other effects prove relevant at the global level as well, and they concern energy consumption, exhaustion of non-renewable energy sources and gas emissions (CO_2 , CFC, CH₄, N₂O, O₃), which ultimately contribute to climate change (Muñiz *et al.*, 2006).

3. URBAN SPATIAL CONFIGURATIONS AND COMMUTING EXTERNALITIES

This section proposes a logical framework for the relationships – to be tested in the next sections – between spatial structure and commuting externalities. The pattern of commutes within an UA is assumed to be affected by its spatial structure, although such a relationship is by no means straightforward. Figure 1 depicts the expected causality links between

spatial structure and socio-demographic characteristics, on the one hand, and per capita CO_2 emissions due to commuting on the other.

The spatial organisation of UAs influences individual travel behaviour through several channels (Ewing and Cervero, 2001; Giuliano and Narayan, 2003; Vance and Hedel, 2008)². Firstly, low-density areas are more difficult to be reached and served efficiently by a pervasive system of public transportation, mainly because of the lack of scale economies (Ellison, 1995). Secondly, the demand for public transportation tends to be lower in a dispersed area, especially where the walking distance from homes to public transport nodes is long enough for cars and motorbikes to prove more competitive. Besides, public means of transport are on average more time-consuming than private ones, especially in low-density areas (Cirilli and Veneri, 2009). Furthermore, the increase in average individual incomes has raised the opportunity cost of time-consuming (public) means of transport.



Figure 1 Flow-chart of factors influencing average per capita CO₂ emissions

Despite a large body of research on this issue, the net effect of compactness – as measured by residential density – on commuting patterns is still ambiguous. On the one hand, the direct effect of density may be a reduction in the distance travelled from home to work (Giuliano, 1989; Banister, 1997), a better environment for walking, cycling and transit services, and a lower level of oil consumption (Newman and Kenworthy, 1989). Moreover, residential density is expected to be positively associated

² A useful review on these topics may be found also in Dieleman *et al.* (2002), Snellen *et al.* (2002).

with the use of public means of transport, which may imply less pollution and congestion externalities (Banister, 1997; Camagni *et al.*, 2002a; 2002b).

On the other hand, densely inhabited areas are usually exposed to a congestion externality (Anas *et al.*, 1998) which tends to be more severe in compact cities, where the bulk of commutes develop radially towards the central business district (CBD). Congested areas are characterised by longer trips for a given distance, which clearly increase the level of polluting emissions. Hence, the analysis of the relationship between density and CO_2 emissions should control for congestion levels.

The environmental implications of a compact model of urban development are difficult to gauge also because the effect of density should be purged of those of other relevant variables that also affect travel behaviour, like functional diversity, accessibility, and the intensity of mass transit use (Cervero and Kockelman, 1997; Cervero and Murakami, 2010). Accessibility, in particular, according to the bid rent theory (Alonso, 1964), plays a fundamental role in shaping the spatial organisation of economic activities and the consequent traffic flows. What is thought to be the effect of density may instead be the effect of accessibility, since the latter reduces both the distance travelled and the duration of commutes, thereby allowing a wider range of choices among the transport means (Cervero and Murakami, 2010; Muñiz and Galindo, 2005). Indeed, the analysis of the relationship between spatial structure and commuting externalities should control for the level of accessibility.

The externalities of commuting are also influenced by other spatial characteristics of UAs, such as the degree of polycentricity and the degree of clustering of the settlements within each area (Tsai, 2001; Veneri, 2010). In smaller areas, the degree of monocentricity often goes hand in hand with the degree of compactness. Both monocentric and polycentric UAs, however, may be compact, albeit to different extents and in different ways. It may be the case, especially for larger (metropolitan) areas, that the most efficient spatial structure is compact *and* polycentric (Carrol, 1977; Edwards, 1977; Haines, 1986). Indeed, as an UA expands, so a single centre becomes more difficult to reach for an increasing number of people (especially those living in the suburbs). A polycentric area, by contrast, may be more accessible, provided it replicates the scale of a smaller and better-organised area – especially if the various sub-centres develop in line with the existing public transportation infrastructure. Hence, the degree of

monocentricity is supposed to be associated with higher CO_2 emissions, especially in large UAs.

Urban size is another important driver of commuting externalities (Cervero and Murakami, 2010 - p. 416), in that journeys tend to be longer in large-scale areas because of a congestion effect. Nevertheless, larger areas are more likely to be endowed with an efficient public transport system, although, on the whole, the distance and the congestion effects are expected to prevail. Urban size is therefore supposed to be associated with higher CO₂ emissions.

Commuting externalities are also affected by functional diversity in land use (Frank and Pivo, 1994; Cervero, 1996; Pouyanne, 2006). A mixed land use tends to reduce the distances travelled by citizens during their circadian cycles, encouraging non-auto commutes for work and other purposes (Cervero, 1996). Hence, a balanced spatial distribution between residential and productive functions is assumed to have a virtuous effect on CO_2 emissions due to commuting. However, the effect of mixed land use on individual travel behaviour also depends on the preference of households – which cannot be taken for granted in the high-income group of workers (Levine, 1998) – to live as close as possible to their workplaces.

Travel behaviours, moreover, also vary according to commuters' socio-economic characteristics, such as age, income and education. Younger workers, for instance, are expected to live in the suburbs because of lower house prices and therefore to take longer and less sustainable trips. The spatial distribution of high-skilled employment appears to be another relevant factor. Given the evidence that more educated people on average travel longer distances to work, a relatively high spatial concentration of high-skilled employment in the CBD – where the bulk of the economic process takes place – may increase the share of co-location between home and workplaces.

Finally, the average age of the housing stock is introduced to capture UAs' recent dynamics, which, at least in Italy, have mainly resulted in scattered development of new settlements, hence requiring longer trips by private cars (Camagni *et al.*, 2002b). Therefore, the higher the proportion of newly built houses, the higher the expected impact of polluting emissions due to commuting.

4. COMMUTING PATTERNS IN THE ITALIAN URBAN AREAS

This section explores the sample of the 111 Italian largest UAs. Due to limited data availability, only home to work trips have been considered, hence ruling out other kinds of journeys (e.g., for educational, consumption or recreational purposes). Likewise, commutes between different UAs have not been taken into account, since the units of analysis are self-contained systems by definition.

The cornerstone of the dataset – which draws on the Italian Population Census – is represented by the number of commuters between any pair of municipalities within each UA (i.e., Local labour system). These trips are disaggregated by transport mode and duration. For each trip from one municipality to another – within the same area –, the point-to-point distance has been computed by the use of Universal Transverse Mercator (UTM) coordinates.

| Mada | Mada decominition | Commu | ters | Distance |
|------|---|-----------|--------|----------|
| Mode | Mode description | abs. val. | % val. | Km |
| | | | | |
| 1 | Train | 98,007 | 1.09 | 13.7 |
| 2 | Tram | 80,100 | 0.89 | 12.3 |
| 3 | Underground | 167,684 | 1.87 | 13.7 |
| 4 | Urban bus or trolley bus | 450,183 | 5.02 | 11.4 |
| 5 | Extraurban bus or coach | 107,566 | 1.20 | 12.2 |
| 6 | School or business bus | 50,940 | 0.57 | 10.8 |
| | Public transport | 954,480 | 10.64 | 12.3 |
| 7 | Private car (driver) | 5,557,294 | 61.93 | 11.0 |
| 8 | Private car (passenger) | 420,680 | 4.69 | 10.1 |
| 9 | Motorbikes or scooters | 624,071 | 6.96 | 9.5 |
| 10 | Bikes, foot or other means of transport | 1,416,359 | 15.78 | 8.4 |
| | Private transport | 8,018,404 | 89.36 | 9.7 |
| | | 8,972,884 | 100.00 | 11.0 |

Table 1: Commuters and point-to-point distances by modes of commuting within the Italian urban areas, 2001

Source: our elaboration on Istat Population Census data, 2001

The choice of the commuting mode is strongly biased towards private means of transport (89.4% over the total number of commuters; Tab. 1). Even ruling out those commuters that travel on foot, by bike or as passengers in a private car, this share is still high (68.9%). Hence, public transportation accounts for only 10% of total commutes and, among all the available options, only urban and trolley buses concentrate a significant share of mobility demand (5%), while rail transport is much less widespread. It is worth noting, however, that longer distances are travelled on average when public means of transport are used (12.3 Km as opposed to 9.7; Tab. 1).

Other relevant insights may be obtained by looking at the average duration of commutes, which strongly varies depending on the transport mode (private vs. public). In fact, around 52% of public transport users spend 30 minutes or longer in their travels, whereas 81.2% of private transport users spend less than 30 minutes (Tab. 2). This discrepancy may have a twofold explanation: on the one hand, commuters prefer public means of transport when they have to travel longer distances. Public means of transport, however, could be relatively less efficient than private ones for a given distance, since the duration of commutes does not necessarily reflect physical distance.

Both factors are expected to come into play. Indeed, when mass transit is used, it turns out that the average distance travelled by commuters increases with the commute duration, while this does not hold for private transport users (see Tab. 2). This finding seems to suggest that public means of transport are relatively less efficient when short distances are travelled, while private transportation is exposed to congestion problems especially in the longer trips.

| Commute duration | Public tr | ansport | Private transport [*] | | | | |
|---------------------|--------------|--------------|--------------------------------|--------------|--|--|--|
| minutes | commuters, % | distance, Km | commuters, % | distance, Km | | | |
| | | | | | | | |
| 0-15 | 11.6 | 8.0 | 50.9 | 7.0 | | | |
| 15-30 | 36.3 | 10.4 | 35.1 | 10.4 | | | |
| 30-60 | 42.0 | 13.2 | 12.6 | 14.5 | | | |
| > 60 | 10.1 | 16.3 | 1.5 | 14.9 | | | |

Table 2: Commuters and average distances by duration of commuting within the Italian urban areas, 2001

^{*} private transport here does not include bikes and othermeans of transport *Source*: our elaboration on Istat Population Census data (2001)

Table 3: Shares of urban commuters that use public means of transport by macro-region and population class, % values, 2001

| UAs' population | North | Centre | South | Islands |
|-----------------|-------|--------|-------|---------|
| | | | | |
| >400,000 | 16.0 | 17.3 | 13.6 | 6.7 |
| 200,000-400,000 | 6.1 | 4.2 | 4.2 | 5.5 |
| 100,000-200,000 | 6.1 | 3.2 | 4.7 | 2.6 |
| <100,000 | 2.3 | 2.2 | 3.1 | 1.1 |

Source: our elaboration on Istat Population Census data (2001)

| Commute | | | | | | | | | |
|----------|-------|----------|-----------|---------|----------------------------|--------|-------|---------|--|
| duration | Pub | lic mean | s of tran | sport | Private means of transport | | | | |
| minutes | North | Centre | South | Islands | North | Centre | South | Islands | |
| | | | | | | | | | |
| 0-15 | 11.2 | 10.4 | 14.8 | 17.1 | 56.3 | 55.4 | 62.5 | 57.7 | |
| 15-30 | 37.2 | 31.6 | 42.2 | 46.4 | 31.3 | 30.1 | 29.3 | 34.0 | |
| 30-60 | 43.0 | 43.8 | 36.7 | 31.7 | 11.3 | 12.6 | 7.3 | 7.6 | |
| > 60 | 8.7 | 14.2 | 6.3 | 4.8 | 1.2 | 1.9 | 0.9 | 0.7 | |
| | | | | | | | | | |
| total | 100.0 | 100.0 | 100.0 | 100.0 | 100.0 | 100.0 | 100.0 | 100.0 | |

Table 4: Commute duration by mode of transport and macro-region, % values, 2001

Source: our elaboration on Istat Population Census data (2001)

On looking at the share of public transport users, it is found that it increases with population size, whatever the macro-region an UA belongs to (Tab. 3). On average, Northern UAs show higher shares of public transport users than the others (especially those in the South and in the Islands). This gap is clearly associated with the high degree of heterogeneity of the Italian macro-regions in terms of average income, institutional quality and infrastructural endowment.

The average duration of intra-urban commutes has been also analysed with respect to the macro-region (Tab. 4) of the UAs, distinguishing in each case between private motorised and public modes of transport. From Table 4 it is worth noting that in Southern and Islands' UAs, on average, the share of public transport users that travel for less than 30 minutes is higher than in Central and Northern Italy.

5. DATA AND VARIABLES

5.1 The dependent variable

The environmental impact indicator proposed in what follows builds upon the methodology of Camagni *et al.* (2002b), but introduces two novelties with respect to the weighting technique of commutes. First of all, transport modes (e.g. car, train, underground, bus, etc.) are weighted on the basis of the level of CO_2 emissions per passenger per kilometre, as estimated by Amici della Terra (2005; Tab. 5).

Table 5 Per passenger per kilometre CO₂ emissions by transport mode

| Means of transport | CO ₂ (grams/km) |
|-----------------------|----------------------------|
| Train | 35 |
| Tram | 32 |
| Underground | 21.3 |
| Urban Bus | 72 |
| Extra-urban Bus | 26 |
| School or Company Bus | 31 |
| Car | 105 |
| Motorbike | 80 |
| Bike, on foot, other | 0 |

Source: Amici della Terra (2005)

The weights attached to the different transport modes take into account the amount of CO_2 emissions generated by each category of transport means, depending on the number of passengers that can be accommodated, as well as on the time needed to travel a given distance. Public means of transport, indeed, carry on average many more passengers than a private ones, with a significant saving in energy and a reduction in polluting emissions. Moreover, all public transport means (especially rail transportation) make a smaller contribution to congestion.

Secondly, the longer the commute, the larger its impact in terms of both (air and noise) pollution and congestion. In this case, however, commutes are weighted according to their distances, rather than their duration, thus obtaining an absolute measure of CO_2 emissions generated by each UA's "typical commuter". Indeed, the duration of commutes is a somewhat ambiguous measure, since a longer duration may reflect either a longer distance travelled or a less efficient mode of transport given a certain distance; all the more so as the time efficiency of the different categories of transport means is already embedded in their respective weights.

 CO_2 emissions per commuter per kilometre are calculated for each UA as in (1). All commutes – disaggregated by transport modes – between any couple of municipalities within each UA have been re-aggregated by weighting with the transport mode as in Amici della Terra (2005), and with the physical distance between that pair of municipalities:

(1)
$$pc_{-}CO_{2} = \sum_{i} \sum_{j} \sum_{k} f_{ijk} w_{k} d_{ij} / \sum_{i} \sum_{j} \sum_{k} f_{ijk}, \qquad k=1,2...10;$$

where f_{ijk} are the commuters that use the *k*-th means of transport and travel from municipality *i* to municipality *j* within a given UA; w_k is the weight attached to the *k*-th means of transport – as shown in Table 5 – and d_{ij} is the distance between the *i*-th and the *j*-th municipality.

Table 6 shows some descriptive statistics, both at the sample and subsample levels. The overall variation range is between 241 (Castellammare di Stabia) and 2 475 (Rome). On average, the lowest levels of emissions are found in Northern UAs and the highest in those of the Islands and Central Italy. One may also note that the indicator takes higher values in the smalland medium-scale areas.

| Urban areas | Obs | Mean | Std. Dev. | Min | Max |
|-------------------|-----|--------|-----------|-------|---------|
| | | | | | |
| Italy | 111 | 948.8 | 385.2 | 240.8 | 2475.0 |
| | | | | | |
| North | 28 | 685.6 | 164.7 | 452.8 | 1,078.6 |
| Centre | 39 | 1081.9 | 386.6 | 553.7 | 2,475.0 |
| South | 28 | 911.7 | 415.5 | 240.8 | 1,691.4 |
| Islands | 16 | 1150.2 | 358.9 | 385.5 | 1,794.3 |
| | | | | | |
| > 400,000 | 17 | 954.7 | 439.8 | 452.8 | 2475.0 |
| 200,000 - 400,000 | 26 | 777.3 | 276.8 | 288.6 | 1256.9 |
| 100,000 - 200,000 | 47 | 991.1 | 358.8 | 240.8 | 1859.9 |
| < 100,000 | 21 | 1061.8 | 462.3 | 385.5 | 1755.0 |

Table 6 CO_2 per commuter: summary statistics over the whole urban system and over subsamples by macro-region and population class, 2001

Source: our calculations on Istat Population Census data, 2001





Source: our elaboration on Istat Population Census data (2001)

In Figure 2 the Italian UAs are mapped by quartile classes of CO_2 emissions per commuter. Overall, the worst performing UAs are found in Central Italy (especially in Lazio and in some areas of Tuscany and Emilia

Romagna), as well as in the Islands and in some areas of Puglia. In regard to metropolitan areas (i.e., UAs with more than 400 000 inhabitants in 2001), intra-urban commutes tend to have a higher environmental impact in Rome and a few Southern UAs like Taranto, Cagliari, Palermo and Catania, while Lombardy's UAs – including Milan – fare relatively better.

5.2 The independent variables

Different patterns of intra-urban commutes may generate different levels of environmental impacts. These patterns, in their turn, may vary depending on how UAs are spatially organised. This section, therefore, presents some variables that capture the spatial configuration of the UAs in the sample and that are thought to influence per commuter CO_2 emissions, according to the model depicted in Figure 1. In Table 7 all the relevant variables are defined and reported with their means and standard deviations. All of them have been computed drawing on the 2001 Istat Population Censuses, Istat Industry and Services Census and Isfort.

Urban form is measured from both an intensity-based and a spatial structure-based perspective. In the former, gross residential density (density) approximates the UA's degree of compactness (i.e., the higher the density, the more compact the area). In the latter, three variables are introduced in order to describe the UA's spatial configuration. The degree of urban monocentricity is measured by the share of employment concentrated in the central municipality. Then, two concentration indexes à la Gini are computed: one (gini_area_empl) is used to assess whether urban employment is evenly distributed from a spatial point of view or concentrated in some municipalities (Tsai, 2005). The other (gini_pop_empl) works as a proxy for functional diversity at the urban level, because it takes a value of zero when population distribution reflects employment distribution from a spatial point of view, implying that residential and productive functions match. The spatial structure of UAs also depends on their size, which is measured here as the employment level in each UA.

As regards the social characteristics of commuters, the empirical analysis takes account of the age of population – measured with a standard index of demographic structure (see Table 7) – and the spatial distribution

of high-skilled workers – measured as the share of graduates in the central municipality over the total number of each UA's graduates.

The accessibility of UAs is also controlled for, by taking into account their infrastructural endowment as approximated by the Isfort accessibility index (Ministero dei Trasporti e delle Infrastrutture, 2005). Finally, account is taken of the recent dynamics of the UAs by looking at the proportion of houses built after 1982 over the total number of houses in 2001 (*house_age*). This variable is expected to assume higher values in those UAs where rapid urbanisation processes have taken place in recent decades.

| | | | | Descr | iptive |
|------------------|---|---|--------------|---------|---------|
| | Conceptual | | | stati | stics |
| Variable name | meaning | Variable description | Data source | | std. |
| | Q | | | mean | dev. |
| | | average daily CO2 emissions per | Istat, Amici | | |
| pc_CO2 | CO2 emissions | commuter. | della Terra | 948.84 | 385.23 |
| | PM10 and NOx | average daily PM10 and NOx | Istat, Amici | | |
| pc_PM10_NOx | emissions | emissions per commuter | della Terra | 52.69 | 21.53 |
| density | compactness of the built environment | residential density (population over total area) in 2001 | Istat | 466.58 | 520.84 |
| pivot_empl_share | degree of monocentricity | central municipality's share of total employment ^b of the UA in 2001 | Istat | 0.65 | 0.18 |
| gini_area_empl | employment concentration | sum, for each municipality within a UA, of the differences in absolute value between the area and the employment ^b shares of that municipality over the whole UA in 2001 | Istat | 0.40 | 0.17 |
| gini_pop_empl | functional diversity | sum, for each municipality within a UA, of the differences in absolute value between the population and the employment ^b shares of that municipality over the whole UA in 2001 | Istat | 0.12 | 0.06 |
| house_age | recent urban dynamics | proportion of houses built after 1982 over the total number of houses in 2001 | Istat | 0.21 | 0.06 |
| empl | size of the UA | number of employed people in 2001 | Istat | 113,322 | 206,258 |
| demo_str | age of population | population structure index: people aged between 40 and 65 over people aged between 15 and 39 | Istat | 60.49 | 8.31 |
| grad_distr | spatial distribution of human capital | central municipality's share of total graduates of the UA in 2001 | Istat | 0.10 | 0.03 |
| accessibility | accessibility | endowment of transport infrastructures | Isfort | 61.22 | 12.85 |
| accidents | congestion | share of traffic accidents over the total number of commuters in 2001 | Istat | 0.05 | 0.02 |

Table 7 Variables descriptions, sources and statistics for 111 Italian urban areas

^a according to Istat classification

^b people that work in the city but do not necessarily reside in it

| | pc_CO2 | NOx_PM10 | density | pivot_emp | gini_areal | gini_pop | house | empl | demo | grad | access. | accid. | public_sh |
|------------------|--------|----------|---------|-----------|------------|----------|-------|-------|-------|-------|---------|--------|-----------|
| | | | | | | | | | | | | | |
| pc_CO2 | 1 | | | | | | | | | | | | |
| pc_NOx_PM10 | 1.00 | 1 | | | | | | | | | | | |
| density | -0.59 | -0.57 | 1 | | | | | | | | | | |
| pivot_empl_share | 0.50 | 0.51 | -0.30 | 1 | | | | | | | | | |
| gini_area_empl | -0.26 | -0.27 | -0.01 | -0.14 | 1 | | | | | | | | |
| gini_pop_empl | -0.12 | -0.12 | -0.14 | -0.26 | 0.57 | 1 | | | | | | | |
| house_age | 0.14 | 0.13 | -0.18 | -0.20 | -0.18 | 0.17 | 1 | | | | | | |
| empl | -0.04 | -0.04 | 0.47 | -0.34 | 0.38 | 0.22 | -0.21 | 1 | | | | | |
| demo_str | 0.06 | 0.03 | -0.07 | 0.13 | 0.20 | -0.13 | -0.68 | 0.31 | 1 | | | | |
| grad_distr | -0.24 | -0.26 | 0.06 | -0.50 | 0.36 | 0.56 | 0.20 | 0.41 | -0.05 | 1 | | | |
| accessibility | -0.28 | -0.29 | 0.40 | -0.29 | 0.07 | -0.10 | -0.32 | 0.51 | 0.45 | 0.16 | 1 | | |
| accidents | 0.11 | 0.07 | 0.08 | 0.24 | -0.08 | -0.27 | -0.28 | -0.03 | 0.47 | -0.35 | 0.17 | 1 | |
| public_share | -0.03 | 0.01 | 0.38 | 0.11 | 0.31 | 0.13 | -0.47 | 0.66 | 0.26 | 0.09 | 0.22 | -0.09 | 1 |

Table 8 Bivariate correlations among independent variables

Source: our on Istat Census data (1981, 2001)

Before the relationship between the spatial configuration of UAs and their per commuter impact in terms of CO₂ emissions is tested empirically, some summary statistics are presented (Tab. 7), along with the correlation structure among all the relevant variables (Tab. 8). Table 8 suggests that the impact of commuting tends to be lower in more compact and monocentric UAs, where public transportation proves relatively more competitive. The share of mass transit users is likely to be higher in more compact UAs, where employment is more spatially concentrated and where urban dynamics have been less intense in the last two decades covered by the analysis. In addition, the intensity of the urbanisation process in recent decades is associated with a lower share of public transport users (Tab. 8). This may indicate that new settlements have been established in discontinuous manner and not necessarily in line with the pre-existing public transport infrastructure, as already highlighted by Camagni, Gibelli and Rigamonti (2002b). In newly urbanised settlements, commuters tend to use private motorised means of transport, but they also travel shorter distances to their workplaces on average, because they tend to find accommodation close to the latter. Overall, the negative environmental effect induced by the transport mode choice is likely to be only partly off-set by the positive distance effect.

6. EMPIRICAL RESULTS AND INTERPRETATIONS

In the econometric analysis, the environmental impact indicator – as computed in Section 5.1 – was regressed on the main variables describing UAs' spatial configurations and their functional diversity, as well as on other variables meant to control for UAs' infrastructural accessibility and their recent dynamics, according to the model depicted in Figure 1.

On investigating the effects of spatial structure on commuting patterns, indeed, several studies have sought to control for endogeneity and simultaneity issues, especially with respect to the effect of density on the number of vehicle miles travelled (VMT) (Bhat and Guo, 2007; Brownstone and Golob, 2009). In fact, individual decisions about residential and job locations may be influenced by the preference for a given mode of transport and a shorter commute duration. In addition, the UAs' spatial structure may be affected by the quali-quantitative endowment of public transport infrastructures.

In our analysis, all the characteristics of spatial configuration were treated as potential sources of endogeneity biases. Hence, in addition to OLS, we carried out a two-stage least square (2SLS) estimation where the four main spatial structure variables – those related to density, concentration, degree of monocentricity and functional diversity – were instrumented by the corresponding variables referred to 1951 data. As shown in Table 9, the estimates obtained with the OLS and the 2SLS methods gave rise to very similar results. In particular, almost all the main regressors related to urban spatial structure are significant at either a 99% or a 95% confidence level. The only two variables that do not seem to play a significant role in the explanation of CO₂ emissions are those used to control for UAs' functional diversity and the age of their housing stock. As regards the model's goodness of fit, more than eighty percent of total variance is accounted for, which seems a good result for a cross-section analysis.

The main finding is that densely inhabited UAs appear to be more sustainable in terms of per commuter CO_2 emissions, even after controlling for accessibility and socio-demographics. Moreover, monocentric UAs – as well as UAs with an even spatial distribution of employment – generate higher levels of polluting emissions. This finding may suggest a certain virtuosity of a polycentric spatial configuration – i.e., decentralised and concentrated – consistently with the recent findings by Veneri (2010) on Italian metropolitan areas. Moreover, the larger the city, the higher the level of per commuter CO_2 emissions, since in larger UAs commuters travel longer distances, thereby offsetting the beneficial effect due to a more intense use of public means of transport.

The accessibility of the UAs is associated, as expected, with lower levels of CO_2 emissions, since in more accessible areas commuters are able to reach their workplaces more rapidly and with a wider set of options among transport means. On the other hand, traffic congestion – approximated by the share of yearly accidents over the total number of commuters – implies longer journeys and therefore generates higher levels of emissions.

Commuters' socio-demographic characteristics play a role as well. In particular, a high concentration of skilled workers in the central municipality entails more environmental externalities, consistently with the assumption that more educated people tend to travel longer distances (Schwanen *et al.* 2001). More specifically, graduates are likely to obtain

better jobs (and better pay), thus developing a preference for larger houses in the outskirts, which in turn implies longer commutes to the CBD.

| vomabla | | | | | IV. | | IV. | |
|------------------|--------------|-----|--------------|-----|--------------|-----|--------------|-----|
| Variable | OLS | | OLS | | 1V | | 1V | |
| (intercept) | -0.505 | | -3.294 | *** | -0.437 | | -3.251 | *** |
| | 0.633 | | 0.006 | | 0.705 | | 0.010 | |
| density | -0.464 | *** | -0.483 | *** | -0.480 | *** | -0.486 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| pivot_empl_share | 0.491 | *** | 0.433 | *** | 0.389 | *** | 0.406 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| gini_area_empl | -0.283 | *** | -0.265 | *** | -0.331 | *** | -0.342 | *** |
| - | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| gini_pop_empl | -0.018 | | -0.047 | | -0.005 | | 0.056 | |
| | 0.635 | | 0.173 | | 0.961 | | 0.488 | |
| house_age | 0.128 | | 0.173 | ** | 0.088 | | 0.148 | * |
| | 0.144 | | 0.026 | | 0.343 | | 0.081 | |
| empl | 0.411 | *** | 0.400 | *** | 0.422 | *** | 0.407 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| demo_str | -0.605 | ** | 0.311 | | -0.599 | ** | 0.313 | |
| | 0.027 | | 0.286 | | 0.044 | | 0.314 | |
| grad_distr | -0.420 | *** | -0.299 | ** | -0.469 | ** | -0.469 | *** |
| | 0.006 | | 0.029 | | 0.015 | | 0.009 | |
| accessibility | -0.273 | ** | 0.142 | | -0.321 | ** | 0.135 | |
| | 0.021 | | 0.252 | | 0.014 | | 0.307 | |
| accidents | 0.214 | *** | 0.097 | | 0.227 | *** | 0.107 | |
| | 0.005 | | 0.152 | | 0.004 | | 0.136 | |
| x_coord | | | 0.000 | | | | 0.000 | |
| | | | 0.372 | | | | 0.564 | |
| y_coord | | | 0.000 | *** | | | 0.000 | *** |
| | | | 0.000 | | | | 0.000 | |
| | | | | | | | | |
| n. obs. | 111 | | 111 | | 111 | | 111 | |
| Adj. R-squared | 0.818 | | 0.863 | | 0.812 | | 0.848 | |
| F-statistic | 50.51(0.000) | | 58.54(0.000) | | 45.38(0.000) | | 50.4(0.000) | |
| mean VIF | 2.29 | | 2.79 | | | | | |
| Breusch-Pagan | 5.55(0.852) | | 17.25(0.141) | | 5.55(0.852) | | 17.25(0.141) | |
| Reset | 2.27(0.109) | | 1.43(0.245) | | 2.27(0.109) | | 1.43(0.245) | |
| Moran's I | 3.01(0.001) | | 1.20(0.115) | | 2.42(0.008) | | 0.46(0.323) | |

Table 9 Estimation results: dependent variable: pc_CO2

*** statistically significant at 99%; ** statistically significant at 95%; * statistically significant at 90%.

| variable | OLS | | OLS | | IV | | IV | |
|------------------|--------------|-----|--------------|-----|--------------|-----|--------------|-----|
| | | | | | | | | |
| (intercept) | 6.723 | *** | 4.000 | *** | 6.946 | *** | 4.037 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.001 | |
| density | -0.426 | *** | -0.444 | *** | -0.448 | *** | -0.455 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| pivot_empl_share | 0.478 | *** | 0.422 | *** | 0.360 | *** | 0.374 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| gini_area_empl | -0.269 | *** | -0.252 | *** | -0.300 | *** | -0.309 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| gini_pop_empl | -0.013 | | -0.040 | | -0.038 | | 0.017 | |
| | 0.726 | | 0.213 | | 0.665 | | 0.816 | |
| house_Age | 0.059 | | 0.098 | | 0.014 | | 0.067 | |
| | 0.467 | | 0.174 | | 0.871 | | 0.381 | |
| empl | 0.391 | *** | 0.380 | *** | 0.404 | *** | 0.389 | *** |
| | 0.000 | | 0.000 | | 0.000 | | 0.000 | |
| demo_str | -0.749 | *** | 0.116 | | -0.783 | *** | 0.114 | |
| | 0.004 | | 0.669 | | 0.005 | | 0.687 | |
| grad_distr | -0.425 | *** | -0.306 | ** | -0.418 | ** | -0.405 | ** |
| | 0.003 | | 0.017 | | 0.020 | | 0.013 | |
| accessibility | -0.250 | ** | 0.132 | | -0.316 | *** | 0.121 | |
| | 0.023 | | 0.253 | | 0.010 | | 0.316 | |
| accidents | 0.164 | ** | 0.057 | | 0.180 | ** | 0.068 | |
| | 0.019 | | 0.367 | | 0.013 | | 0.296 | |
| x_coord | | | 0.000 | | | | 0.001 | |
| | | | 0.243 | | | | 0.305 | |
| y_coord | | | -0.006 | *** | | | -0.005 | *** |
| | | | 0.000 | | | | 0.000 | |
| n. obs. | 111 | | 111 | | 111 | | 111 | |
| Adi. R-squared | 0.816 | | 0.861 | | 0.807 | | 0.853 | |
| F-statistic | 49.85(0.000) | | 57.59(0.000) | | 43.98(0.000) | | 52.07(0.000) | |
| mean VIF | 2.29 | | 2.79 | | , .() | | | |
| Breusch-Pagan | 5.49(0.856) | | 19.29(0.082) | | 5.49(0.856) | | 19.27(0.082) | |
| Reset | 2.29(0.106) | | 1.42(0.246) | | 2.29(0.106) | | 1.42(0.246) | |
| Moran's I | 2.64(0.004) | | 1.37(0.085) | | 2.19(0.014) | | 0.56(0.286) | |

Table 10 Estimation results: dependent variable: pc_PM10_NOx

*** statistically significant at 99%; ** statisticalli significant at 95%; * statistically significant at 90%.

The relationship shown in Figure 1 was also tested with respect to pollutant emissions other than CO_2 . In particular, indicators of environmental impact in terms of per commuter PM_{10} and NO_x emissions were alternatively adopted as dependent variables in the estimation framework. These indicators were computed in a similar way to the CO_2 impact indicator (1), using, however, specific weights for the commuting

modes as in Amici della Terra and Enea (2003 - p. 52). Again the results were perfectly consistent with those obtained with CO₂ weights – in terms of sign and size of the coefficients – with both the OLS and the IV estimations (Tab. 10).

Standard diagnostics were carried out to ascertain whether the underlying classic assumptions of the OLS model had been violated. Problems of either heteroskedasticity and multi-collinearity did not seem to arise (see the Breusch-Pagan tests and VIF statistics in Tables 9 and 10). In regard to the possible spatial autocorrelation of residuals, the Moran's *I* statistics showed that in the baseline specifications of the models (CO₂ and NOx & PM₁₀) the residuals were not spatially independent, which could imply biases in the estimated coefficients.

This spatial correlation problem may have been due to the omission of relevant variables, to unobserved spatial heterogeneity, or to a spatial dependence problem. The Reset test reported at the bottom of Tables 9 and 10 ruled out a problem of omitted variables, while spatial dependence did not appear to be relevant in a cross section of UAs, where only intra-urban commutes were taken into account (i.e., a spill-over effect was unlikely). As a result, the autocorrelation of residuals may have been caused by spatial unobserved heterogeneity, which can be corrected by including latitude and longitude coordinates in the regression. As shown by the Moran's *I* statistics, this correction was able to clean off the spatial autocorrelation bias, and this result was robust to the use of different spatial weighting matrixes.³ This finding confirms that commuting behaviours also vary according to the UAs' geographical location, and especially on their longitude (Northern vs. Southern areas).

7. CONCLUDING REMARKS

Cross-section analysis on the 111 largest Italian UAs has corroborated this paper's hypothesis that smaller, more compact, and less monocentric UAs are associated with lower levels of per commuter CO_2 emissions. Population density proves to be one of the main determinants of per commuter emissions – even after controlling for UA accessibility –

³ Moran's *I*s reported in Tables 9-10 were computed using four-nearest-neighbours binary standardised matrixes. The results are consistent with other spatial matrixes, computed on the base of the inverse of distance among UAs, applying different thresholds of minimum distance.

consistently with the findings of important contributions to this field of research in the European literature (Muñiz and Galindo, 2005). The results are also consistent with the assumption that polycentric UAs represent an efficient spatial configuration in terms of per commuter pollutant emissions. Commuters' socio-demographic characteristics also play a role. In particular, travel behaviours seem to be affected by the spatial distribution of the most educated workers and, ultimately, by the population age structure, on which housing preferences also depend. These results prove quite robust after a number of checks are carried out with respect to the estimation technique.

Reducing the negative externalities generated by transport and vehicle use is certainly an important policy goal. These externalities, which mainly consist in traffic congestion and polluting emissions, can be dealt with by urban and regional planners in different ways, also depending on the cultural, institutional and economic background being considered. In the US literature, some scholars suggest that the use of Pigouvian taxes is the most efficient way to curb traffic externalities (Brownstone and Golob, 2009). However, in the Italian – and, to some extent, European – case, where oil taxes and parking tariffs are already high, it would hardly be effective (and politically acceptable) to adopt only these policy tools. The analysis developed in this paper, indeed, confirms that spatial planning - by influencing the functional and morphological organisation of UAs - can be effective in reducing CO₂ and other polluting emissions due to commuting. Hence, the design of new settlements on the basis of a more sustainable pattern of vehicle use represents a fundamental policy strategy, albeit one bound to display its beneficial effects much more slowly than other tools (e.g., taxes). Besides environmental sustainability, efficient urban planning could also enhance the return on public investment in transport infrastructures.

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